PLANT COMMUNITY ANALYSIS AND ECOLOGY OF AFROMONTANE AND TRANSITIONAL RAINFOREST VEGETATION OF SOUTHWESTERN ETHIOPIA

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ABSTRACT: The plant communities of the forests of southwestern Ethiopia were described based on floristic analysis of the data collected between February 1995 and May 1996. Floristic analysis is based on the cover-abundance values of both woody and herbaceous species. Plant community-environment relationship was assessed based on topographic and some soil physical and chemical characteristics. A total of 101 relevés were analysed at altitudes between 1050 and 2550 m a.s.l. (metres above sea level), and a total of 139 species of vascular plants were identified. Nine plant community types were described. Most of these show significant variations for most of the environmental parameters studied. The species in the forest were phytogeographically related to the Afromontane, Guineo-Congolean or Guineo-Congolean linking species.

Key words/phrases: Afromontane, classification, Ethiopia, plant community, transitional rainforest

INTRODUCTION

The forest vegetation cover of Ethiopia has been declining continuously. In the early 1950's, 16% of the land area of Ethiopia was covered with forests (EFAP, 1994). In the early 1980's, the forest coverage was reported at 3.6% and in 1989 it was estimated at only 2.7% (EFAP, 1994). At present, most of the remaining forests of Ethiopia are confined to the south and southwest.

The forests of southwestern Ethiopia are home to various wild animals and contain the gene pool of various endemic and indigenous plants including *Coffea arabica* L., which accounts for more than three-quarters of the world's coffee production, *Aframonum corrorima* (Braun) Jansen, *Scadoxus nutans* (Friis & Bjørnstad) Friis & Nordal, and *Vepris dainellii* (Pich-Serm.) Kokwaro.

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The flora and vegetation of southwestern Ethiopia have been the subject of study by Logan (1946), Chaffey (1979), Friis *et al.* (1982), and Friis (1992). Among these studies, only Friis *et al.* (1982) utilized floristic analysis for the description of the forest vegetation. However, this description neither attempted to identify the plant communities, nor did it explain the environmental relationships of these communities. In the present study, an attempt has been made to address both aspects. Therefore, the objectives of the present study were: (1) to study the floristic composition of the forest vegetation of southwestern Ethiopia with an aim to identify the plant communities; and (2) to understand the ecological relationships between the plant communities and the environmental parameters.

STUDY AREA

The forests investigated in this study stretch between 1050 and 2550 m a.s.l. and include Jiren forest, Yayu forest, Gebre Dima forest, Sele-Anderacha forest, Bonga forest, Sheko forest, and Belete-Gera forest. These forests are located between 6° 53′ to 8°27′ N latitudes and 35°15′ to 36°50′ E longitudes (see Fig. 1). The forest of Sele-Anderacha at 670 kms distance from Addis Ababa is the furthest while Belete-Gera forest is the closest at 380 kms.

The basement of the study area contains Precambrian rocks and Tertiary lavas that lie directly on the crystalline basement. These lava flows were accompanied by, and altered with, the eruption of large amounts of ash and coarser fragmental material forming the trap series (Mohr, 1971).

The soils of the southwest plateau reflect the change of topography and the drop in temperature and increased rainfall associated with higher altitude. Soils developed on the steep lands of escarpments and along ridges are thin and young. Rolling terrain at altitudes between 1200 and 1800 m a.s.l. favours the formation of deep, reddish, well-drained soils which, despite their favourable physical structure and excellent internal drainage, are only moderately fertile. At higher elevations clay illuviation becomes increasingly important and the soils lose their reddish colour to become light brown or yellowish brown. According to the soil classification system by FAO (1974), the soil order of the study area is Nitosol.

The southwestern region of Ethiopia is the wettest with eight rainy months that extend from March to October. According to Daniel Gamachu (1977), the annual rainfall of the highlands in the southwest ranges between 1400 and 2200 mm. However, the rainfall data from the period 1982–1994 from stations nearby show an average annual rainfall between 1600 and 2552 mm. The mean annual temperatures are between 15 to 25° C, the mean daily tempera-

ture minima range from 9.7 to 16.3° C and the corresponding maxima from 20 to 30.4° C.

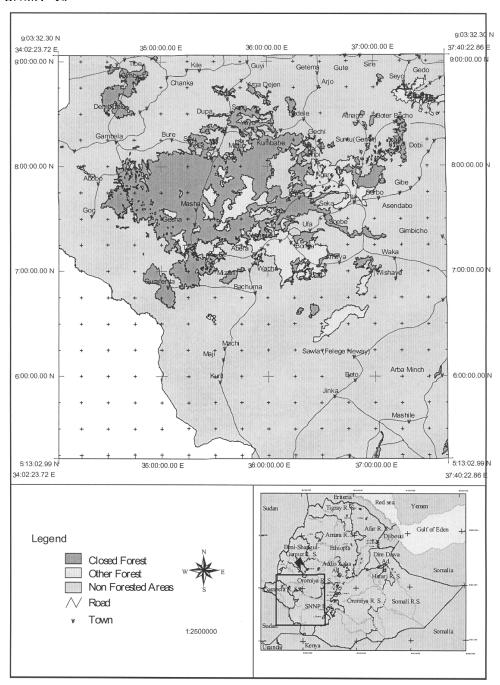


Fig. 1. Study area showing distribution of closed forest.

The vegetation of the study area consists of forests, bush lands, woodlands, and grasslands (Friis *et al.*, 1982). According to Friis (1992), three forest types are recognized in southwestern Ethiopia: lowland dry peripheral semi-evergreen Guineo-Congolean forest, Afromontane rainforest, and Transitional Rainforest. Human impact on the vegetation is severe. Many forested areas have been converted into agricultural land. Some of the forests in the study area may be secondary or belong to different serial stages in the regrowth of the forests.

MATERIALS AND METHODS

Vegetation analysis

A reconnaissance survey was made across the forests of southwestern Ethiopia in March 1995 in order to obtain an impression of the site conditions and physiognomy of the vegetation, and to identify sampling sites. Vegetation data were collected during February 1995, April 1995, December - January 1996 and May 1996.

During sampling, visually checked homogenous representative stands were selected and sample plots of 900 m 2 (30 m \times 30 m) were delimited by relascope and tape. In order to reduce the effect of human interference, the sites selected were, as much as possible, far away from the main road and human settlements. Floristic analysis inside the plot was carried out following the procedure of Braun-Blanquet (see Poore, 1955; 1956; Becking, 1957; Mueller-Dombois and Ellenberg, 1974; Westhoff and Van der Maarel, 1978). All vascular plant species present in the sampling unit were recorded and the non-epiphytic and non-climbing plants were rated according to their cover/abundance using the modified cover-abundance scale of Braun-Blanquet (Van der Maarel, 1979). Floristic analysis of herbaceous species was made on a 2 m x 2 m subplot laid within the larger plot where the vegetation was assumed to be representative.

Identification of plant species was made at the National Herbarium (ETH), Addis Ababa University. Nomenclature follows Cuffodontis (1953–1972), Hedberg and Edwards (1989; 1995) and Edwards *et al.* (1995).

Site description, soil sampling and soil analysis

At each sampling site altitude was measured using "Thommen" altimeter, aspect was measured using "Type 15 SILVA" compass, and slope was measured using a clinometer. Soil samples up to 20 cm depth (topsoil) and 60 cm depth (subsoil) were collected from five points in each sample plot, four at the corners and one at the middle. A composite soil sample weighing about 1.5 kg was taken for further analysis.

Soil analysis was performed in the soil laboratories of Addis Ababa University and Ministry of Agriculture, Addis Ababa, following the procedures outlined in Jou (1978), Chopra and Kanwar (1982), and Dewis and Frietas (1984). The soil samples were analyzed for texture, using the Bouyoucos hydrometer method, pH using a glass electrode pH-meter, and electrical conductivity using conductivity-meter. Available phosphorus was determined following Bray No. 1 method, organic matter according to the Walkely-Black method, and total nitrogen following Kjeldahl method. Exchangeable bases and cation exchange capacity were determined by extracting with ammonium acetate, at pH 7 using flame photometer for sodium and potassium and atomic absorption spectrophotometer for calcium and magnesium.

Data analysis

The cover-abundance values of the species were used as class entities to classify the floristic data using Average-linkage clustering procedure with the program SYNTAX: PROGRAMME NCLAS-Hierarchical clustering by distance optimization (Podani, 1988). In the SYNTAX program, similarity ratio was selected for resemblent coefficient and dissimilarities among the various sample plots were measured using average linkage.

The values of environmental parameters for all sample plots that make up a particular community type were added and averaged for the topsoil and subsoil separately. Aspect was codified according to Zerihun Woldu *et al.* (1989): N = 0; NE = 1; E=2; SE = 3; S = 4; SW = 3.3, W = 2.5; NW = 1.3; Ridge top = 4. Analysis of variance (ANOVA) was performed to detect variation among the community types with respect to any one environmental parameter. Duncan's multiple range test was performed to detect significant differences among the different means of the environmental parameters of each community type. Pearson's product moment correlation coefficient was calculated to evaluate the relationship between the environmental parameters. These statistical analyses were performed with the program STATISTICA.

RESULTS AND DISCUSSION

Floristics

A total of 139 species of vascular plants representing 56 families were recorded from the tree, shrub, and field layers. Of these 41.7% were trees, 10.1% trees/shrubs, 12.9% shrubs, and 35.2% herbs. Seventy-one per cent of the families were dicots, 14.6% monocots, and 14.3% pteridophytes. Seventy-seven per cent of the species were dicots, while 15.1% monocots, and 7.9% pteridophytes. The families with the highest number of species were Euphorbiaceae and Moraceae (10 species each) among the dicots, Poaceae (7 species) among monocots, and Aspleniaceae (3 species) among Pteridophytes. Fifteen species of vascular epiphytes belonging to 9 families were recorded. Four species belonged to Orchidaceae, while 3 species belonged to

Polypodiaceae, and 2 species belonged to Piperaceae. The families Adianthaceae, Amaryllidaceae, Aspleniaceae, Camponviaceae, Lycopodiaceae, and Oleandraceae were represented with one species each. Twelve species of woody climbers (lianas) belonging to 11 families were recorded from the quadrats. Two of the species belonged to the family Celasteraceae/Hippocrateaceae, whereas the rest belonged to Apocynaceae, Araceae, Combretaceae, Menispermaceae, Myrsinaceae, Oleaceae, Phytolaccaceae, Ranunculaceae, Rhamnaceae and Urticaceae. The complete list of the species can be made available from the authors upon request.

Vegetation classification

Nine clusters can be recognized from the SYNTAX output at dissimilarity level above 0.7 (Fig. 2). These clusters were designated as local plant community types and given names after one or two dominating and/or characteristic species, usually a tree and a shrub. A dominant species in this case is a species having a synoptic cover-abundance value (mean frequency \times mean coverabundance, sensu (Van der Maarel $et\ al.$, 1987)) of at least 7, and a characteristic species being a species with synoptic values of 4, 5 or 6 in the type but absent in most of the other community types. Synoptic cover-abundance values for the most important species are shown in Table 1.

The description of the plant community types based on the dominant and characteristic species with their altitudinal distribution is as follows. The cluster numbers in the dendrogram (Fig. 2) correspond to numbers of the community types in the subsequent discussion.

1. Arundinaria alpina type

The *Arundinaria alpina* type is found at altitudes from 2450 to 2550 m. The dominant species of this type is the highland bamboo *Arundinaria alpina* K. Schum. There is no shrub layer in this type. *Dicliptera laxata* C. B. Clarke and *Pilea bambuseti* Engl. are dominant species in the field layer. The characteristic species of the field layer is *Laportea alatipes* Hook. f. Lianas are not present in this community type. Isolated trees of *Syzygium guineense* (Willd.) Dc. and *Schefflera volkensii* (Engl.) Harms are present scattered through the bamboo stand.

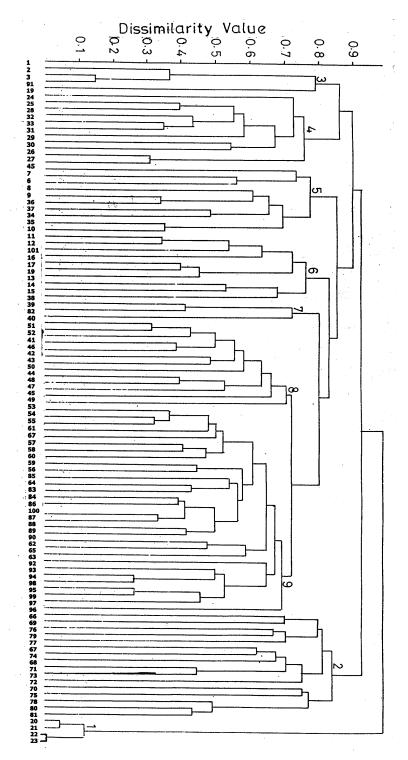


Fig. 2. Dendrogram of the relevè group of sothwestern Ethiopian forest.

Table 1. Synoptic cover-abundance values for species reaching a value of >1.5 in at least one community type; values in bold refer to occurrences as characteristic species.

•				Clus	ter nun	ıber			-
Species	1	2	3	8	5	7	. 9	6	4
Arundinaria alpina	9.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Dicliptera laxata	7.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Pilea bambuseti	7.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Laportea alatipes	5.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Pteris dentata	3.5	0.0	1.3	0.0	0.0	0.0	0.0	4.9	4.4
Coffea arabica	0.0	7.3	0.0	0.0	3.8	0.0	3.7	0.5	0.0
Manilikara butugi	0.0	6.6	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Argomuellera macrophylla	0.0	3.1	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Alchornea laxiflora	0.0	2.1	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Aningeria altissima	0.0	1.6	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Dracaena fragrans	0.0	1.8	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Brillantaisia madagascariensis	0.0	2.6	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Leptaspsis zeylanica	0.0	2.8	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Bersama abyssinica	0.0	0.8	3.5	2.2	1.1	2.7	0.4	1.3	0.3
Canthium oligocarpum	0.0	0.0	1.8	0.2	0.4	0.0	0.0	0.0	0.8
Maytenus gracilipes	0.0	0.0	6.3	0.8	1.7	0.0	0.0	2.1	0.0
Vangueria apiculata	0.0	0.0	3.2	0.0	0.4	0.0	0.0	0.0	0.0
Phaulopsis imbricata	0.0	0.0	2.3	0.9	0.0	0.0	0.0	0.0	0.0
Impatiens hochstetteri	0.0	0.0	7.3	0.0	_ 0.0	0.0	0.0	0.0	3.1
Syzygium guineense	0.0	0.0	7.5	7.1	1.8	2.0	3.1	1.8	2.5
Apodytes dimidiata	0.0	0.0	5.8	4.8	0.0	0.0	0.0	0.5	0.0
Oplismenus compositus	0.0	1.1	0.0	7.5	4.1	0.0	5.9	0.0	0.0
Acanthus eminens	0.0	0.0	0.0	1.7	0.0	0.0	0.0	0.0	0.4
Maytenus obscura	0.0	0.0	0.0	1.7	0.0	0.0	0.9	0.0	0.0
Piper capense	0.0	0.1	0.0	1.9	0.0	0.0	0.0	1.5	0.5
Thelypteris longicuspis	0.0	2.1	0.0	3.4	_0.0	_ 0.0	0.0	0.0	0.0
Dracaena afromontana	0.0	0.8	0.5	5.9	7.1	3.7	1.8	4.2	4.5
Teclea nobilis	0.0	0.0	0.0	2.2	4.8	1.0	2.8	4.1	0.3
Albizia grandibracteata	0.0	0.0	0.2	0.9	2.8	0.0	0.0	0.0	1.6
Celtis africana	0.0	1.3	0.0	0.0	5.1	0.0	0.0	0.0	0.0
Cordia africana	0.0	1.3	0.0	0.0	2.3	0.0	0.0	0.0	0.0
Rothmannia urcelliformis	0.0	2.3	0.0	0.7	3.6	1.7	0.0	2.2	0.0
Pseudechinolaena polystachya	0.0	0.0	1.3	1.8	5.1	8.0	2.2	0.0	0.0
Chionanthes mildbraedii	0.0	0.0	0.0	3.5	3.5	6.0	4.1	3.6	0.0
Rytigynia neglecta	0.0	0.0	1.5	1.4	0.9	2.3	1.6	1.1	1.1
Aningeria adolfi-friederici	0.0	0.0	0.0	1.6	0.9	7.0	0.0	4.2	0.0
Poecilostachyus oplismenoides	0.0	2.1	1.1	1.5	0.0	8.0	1.7	3.4	0.0
Croton macrostachyus	0.0	0.4	1.5	2.3	1.2	6.0	0.0	1.8	0.6
Diospyros abyssinica	0.0	1.0	0.0	0.0	2.8	4.0	0.0	1.7	0.5
Psychotria orophila	0.0	0.0	0.0	3.0	0.8	4.0	0.9	0.0	2.8
Desmodium repandum	0.0	0.6	0.0	5.5	0.0	6.0	0.9	0.0	0.0
Olea welwitschii	0.0	0.0	0.0	2.0	2.5	0.0	5.4	1.5	0.0
Sapium elipticum	0.0	1.2	0.0	0.0	1.2	0.0	2.2	2.7	0.0
Эариан еприсит Aframomum korarima	0.0	0.0	0.0	0.0	0.0	0.0	4.1	2.7	0.0
Vepris daniellii	0.0	0.6	0.0	0.0	0.0	0.0	2.8	0.0	0.0
Sanicula elata	0.0	1.4	0.0	2.4	0.0	0.0	4.0	0.0	0.0
Polyscias fulva	0.0	0.4	0.0	0.4	0.0	0.0	3.0	0.0	
Achyranthus aspera	0.0	1.0	0.0	0.0	0.0	0.0	1.4	0.0	0.0
Ficus sur	0.0	0.1	0.0	1.1					
Justicia schimperiana	0.0	0.0			0.2	1.3	0.8	2.7	0.0
Albizia gummifera	0.0	0.4	0.0 0.0	0.0 0.8	$0.0 \\ 0.1$	0.0	0.0	5.0	0.0
Millettia ferruginea	0.0	1.1	0.0	0.8			0.0	2.9	0.1
Schefflera abyssinica					2.3	0.0	2.7	2.8	0.5
Deinbollia kilimandscharica	0.0	0.0	0.0	2.1	0.0	0.0	2.5	2.7	0.4
	0.0	0.0	0.0	0.0	0.0	0.0	0.0	1.9	1.2
Allophylus abyssinicus Schafflara abyssinica	0.0	0.0	1.3	0.0	0.0	0.0	0.9	5.2	2.6
Schefflera abyssinica	0.0	0.0	0.0	2.1	0.0	0.0	2.5	2.7	0.4
Cyathea manniana Caliniana sanifua a	0.0	0.0	0.0	0.6	0.0	0.0	0.0	0.0	4.1
Galiniera saxifraga	0.0	0.0	0.0	0.7	0.5	0.0	0.5	2.0	5.3
llex mitis	0.0	0.0	0.0	0.0	0.0	0.0	0.0	2.9	5.1
Macaranga capensis	0.0	0.9	1.3	0.0	0.0	0.0	0.0	0.0	3.5
Maytenus arbutifolia Aneilema beninese	0.0	0.0	0.0	0.0	0.0	2.3	0.0	0.0	2.5
	0.0	0.6	0.0	0.0	0.0	0.0	0.0	0.0	2.9
Elatostemma monticolum	0.0	0.0	0.0	2.2	0.0	0.0	0.0	0.0	2.5

2. Manilkara butugi-Coffea arabica type

This community type is distributed at altitudes between 1050 and 1500 m. *Manilkara butugi* Chiov. is the characteristic species in the tree layer and *Coffea arabica* L. is the dominant species in the shrub layer. *Pouteria altissima* (A. Chev) Baehni, *Alchornea laxiflora* (Benth.) Pax & Hoffm., *Argomuellera macrophylla* Pax, *Celtis zenkeri* Engl., and *Dracaena fragrans* (L.) Ker-Gawler are trees and shrubs associated with this type. *Brillantaisia madagascariensis* And. ex Lindau., *Leptaspsis zeylanica* Nees ex Steud., and *Dorstenia sorensenii* Friis are species in the field layer and recorded in this type alone. Other associated species in this type include *Strychnos mitis* S. Moore, *Garcinia buchananii* Bak., *Hallea robrostipulata* (K. Schum.) J. F. Leroy, *Morus mesozygia* Stapf, and *Rungia grandis* T. *Landolphia buchananii* (Hall. F.) Stapf, *Combretum paniculatum* and *Hippocratea africana* (Willd.) Loes. are lianas abundant in this type.

3. Syzygium guineense-Maytenus gracilipes type

This community type is distributed at altitudes between 2280 and 2420 m. *Syzygium guineense* and *Apodytes dimidiata* are the dominant species in the tree layer. *Maytenus gracilipes* (Welw. Ex Oliv.) Exell and *Vangueria apiculata* K. Schum. are the characteristic species in the shrub layer. *Impatiens hochstetteri* Warb. is the dominant species in the field layer. The lianas *Landolphia buchananii* and *Urera hypeselodendron* (A. Rich) Wedd. are well represented in this type.

4. Ilex mitis-Galiniera saxifraga type

The altitudinal distribution of this community type extends from 1950 to 2340 m. a.s.l. The characteristic species of this type are *llex mitis* (L.) Radlk., *Galiniera saxifraga* (Hochst.) Bridson, and the tree fern, *Cyathea manniana* Hook.. Asplenium bugoiense Hieron. and Pteris dentata Forssk. are the characteristic species in the field layer. *Drymaria cordata* (L.) Roem. & Schultes and *Triestemma mauritanum* J. F. Gmelin were herbs recorded from this type only. *Hippocratea africana* and *Landolphia buchananii* are the lianas represented in this type.

5. Celtis africana-Dracaena afromontana type

This type is distributed at altitudes between 1460 and 1960 m. *Celtis africana* Burm. f. is the characteristic species of the tree layer, while *Dracaena afromontana* Mildbr. is the dominant species in the shrub layer. The grass *Pseudechinolaena polystachya* (Kunth) Stapf is the characteristic species in the field layer. *Ritchia albersii* Gilg. was recorded from this community type only. The lianas most abundant in this type are *Landolphia buchananii*, *Combretum paniculatum* Vent., *Hippocratea goetzei*, and *H. africana*.

6. Allophylus abyssinicus-Justicia schimperiana type

This community type is distributed at altitudes between 1760 and 2030 m. *Allophylus abyssinicus* (Hochst.) Radlk. and *Pouteria adolfi-friederici* are the characteristic species in the tree layer. *Justicia schimperiana* is the characteristic

shrub. The dominant species of the field layer is *Pteris dentata*. Lianas and woody climbers represented in this type are *Hippocratea goetzei*, *H. africana*, *Tiliacora troupinii* Cuffod., *Combretum paniculatum*, and *Jasminum abyssinicum* Hochst. ex DC.

7. Pouteria adolfi-friederici -Chionanthes mildbraedii type

This community type is distributed at altitudes between 1880 and 1940 m. *Pouteria adolfi-friederici* (Engl.) Baehni is the dominant species of the tree layer. *Croton macrostachyus* Del. and *Diospyros abyssinica* (Hiern) F. White are the characteristic species in the tree layer. *Chionanthes mildbraedii* (Gilg & Schellenb.) Stearn is the characteristic species in the shrub layer. *Pseudechinolaena polystachya* and *Poecilostachyus oplismenoides* (Hack.) W. D. Clayton are dominant in the field layer and *Desmodium repandum* (Vahl) Dc. is the characteristic species in the same layer. *Landolphia buchananii* and *Hippocratea goetzei* are the lianas most abundant here.

8. Syzygium guineense-Thelypteris longicuspis type

This type is distributed at altitudes from 1845 to 2240 m. The dominant species in the tree layer is *Syzygium guineense*. *Apodytes dimidiata* is the characteristic species. *Oplismenus compositus* is the dominant grass in the field layer and *Desmodium repandum* is the characteristic herb. *Landolphia buchananii* and *Combretum paniculatum* are the lianas abundant in this type.

9. Olea welwitschii-Aframomum corrorima type

This community type is distributed between 1675 and 2010 m. a.s.l. This community type is characterized by *Olea welwitchii* (Knobl.) Gilg. & Schellenb. in the tree layer. *Schefflera abyssinica* (Hochst. ex A. Rich.) Harms, *Polyscias fulva* (Hiern) Harms and *Vepris daniellii* (Pichi-Sermolli) Kokwaro are the associated trees. *Chionanthes mildbraedii* is the characteristic shrub. *Oplismenus compositus* (L.) P. Beauv., *Sanicula elata* Ham. Ex Don, and *Aframonum corrarima* (Per.) Engler are characteristic herbs in the field layer. *Landolphia buchananii*, *Hippocratea africana*, *H. goetzei*, and *Combretum paniculatum* are common lianas prersent in this type.

Community-environment relationship

The community types identified from the classification output showed significant differences with respect to all the environmental parameters in the topsoil. However, significant differences were observed in slope, aspect, exchangeable sodium, and exchangeable potassium, cation exchange capacity, and available phosphorus in the subsoil. The values of the various environmental parameters averaged for each plant community type are presented in Table 2.

Comm.	Comm. Altitude Slope Type (m) (°)	Slope 1	Exposure	Sand (%)		Silt (%)		Clay (%)		Hd		Conductivity (mmhos/cm)	vity cm)	Organic matter(%)	υ (9
1				TS	SS	TS	SS		SS	TS	SS	TS	SS	TS	8
1	2505 ⁴ ±28.86	8.3 ^{ns} ±2.17	2.0 ^{rs} ±0.09	71.8 ^{ad} ±3.4	58.0° ±3.72	20.4 ^{ab} +2.52	25.9abc ±1.1	7.9ac +2.59	16.0 ^{ab} ±2.89	4.1 ^a ±0.09	4.4ª ±0.29	1.2° ±0.02	0.67a ±0.05	20.1 ^a ±1.08	14.2° ±0.95
7	1338 ^f ±61.15	15.5 ^{ns} ±1.88	2.5 ^{ns} ±0.12	43.6 ^{bc} ±3.66	34.5 ^b ±4.2	28.4 ^{ab} ±1.89	30.4 ^{abc} ±2.03	28.1 ^b ±2.48	34.5 ^{ab} ±4.08	5.84 ^b ±0.12	5.57° ±0.18	0.46 ^b ±0.02	0.13 ^b ±0.01	10.0 ^b ±0.53	4.5 ^d ±0.45
6	2378ª ±32.75	13.0°s ±4.69	3.0 ^{ns} ±0.05	62 ^{abd} +2.35	47.3ab ±3.55	27.3 ^{ab} ±3.68	31.9abc ±3.51	10.8 ^{abc} ±1.72	20.9ab ±3.28	5.32 ^b ±0.05	5.09 ^{ab} ±0.17	0.32bc ±0.01	0.24 ^b ±0.03	12.8°d ±0.79	8.6 ^{bc} ±0.79
4	2152 ^b ±44.95	9.6 ^{ns} ±1.87	2.5 ^{ns} ±0.08	70.4 ^d ±3.56	60.3ª ±4.59	18.3 ^b ±2.18	19.9ª ±2.68	11.4° +2.22	19.8 ^a +2.28	4.2ª ±0.08	4.43° ±0.08	0.26 ^{bc} ±0.02	0.2 ^b ±0.01	14.5° ±0.99	7.6°
ω '	1729° ±63.0	12.4 ^{ns} ±1.88	2.8 ^{ns} ±0.09	51.2 ^b ±3.37	37.7° ±3.74	27.8° ±2.34	26.6 ^{abc} ±3.36	21.0abc ±4.13	35.6 ^{ab} ±3.33	5.89 ^b ±0.09	5.75 ^b	0.44 ^{bc} ±0.02	0.2 ^b ±0.01	15.2 ^{bd} ±0.54	4.8 ^{bd} ±0.29
9	1880cde ±33.97	11.0°s ±3.71	1.6°s ±0.15	35.8° ±4.66	33.0 ^b ±3.51	28.0° 142.73	31.3abc ±2.89	26.1 ^{abc} ±2.72	35.7ab ±4.91	4.76° ±0.15	4.77a ±0.08	0.48 ^{bc} ±0.03	0.26 ^b ±0.01	9.5 ^{bd} ±0.53	5.7 ^{td} ±0.54
7	1910 ^{bcde} ±17.32	6.7 ^{ns} ±1.73	4.0 ^{ns} ±0.11	52.0abc ±4.92	25.4 ^b	34.3ª ±2.85	45.6° ±4.69	13.6 ^{abc} ±1.31	29.0 ^{ab} ±6.64	5.82 ^b ±0.11	5.98b ±0.41	0.54 ^{bc} ±0.01	0.14 ^b ±0.02	8.6 ^{bod} ±0.57	3.6 ^{bcd} ±0.31
∞	2021 ^d 36.28	12.5°s ±1.15	2.4 ^{ns} ±0.11	46.4 ^{bc} ±2.46	28.4 ^b ±1.49	31.2° ±1.56	32.3*c ±1.56	22.5abc ±1.78	39.6 ^b ±1.68	5.66° ±0.11	5.4 ^b ±0.07	0.38 ^{bc} ±0.02	0.26 ^b ±0.02	8.6 ^b	4.8 ^d ±0.19
6	1860° ±18.29	10.1 ^{ns} ±1.15	2.1 ^{ns} ±0.08	48.3b ±1.19	32.2 ^b ±0.93	35.0° ±1.89	40.1 ^{bc} ±2.46	16.7ac	27.6 ^{ab} +2.44	5.76 ^b ±0.08	5.44 ^b ±0.07	0.26° ±0.01	0.22 ^b ±0.01	8.6 ^b ±0.21	4.0 ^d ±0.13

able 2 (contd.)

mge m)	88	34.2 ^{ns} ±6.51	27.6°s ±3.28	41.1 ^{ns} ±2.83	38.57s ±2.06	39.1 ^{ns} ±5.18	29.8 ^{ns} ±2.42	28.3 ^{ns} ±2.45	32.0°s ±1.39	27.4 ^{ns} ±1.14
Cation Exchange Capacity (meq/100gm)	TS	45.9 ^{ab} +2.53	31.6ª	53.4 ^b ±6.77	46.8 ^b ±3.46	49.0 ^b ±5.23	36.6 ^{ab} +2.39	38.9ab ±2.53	39.5 ^{ab} ±1.76	30.9ª ±0.86
n Se	88	0.71ª ±0.05	2.67a ±0.56	2.65 ^{ab} ±0.63	0.98a ±0.37	5.14 ^b ±0.88	1.64ª ±0.21	2.78 ^{ab} ±0.63	3.18 ^a ±0.67	1.94ª ±0.19
Exchangeable Magnesium (meq/100gm)	TS	1.23 ^{ab} ±0.21	3.66 ^{bc} ±0.56	5.46 ^{bc} ±1.13	1.26ª ±0.28	5.54° ±0.93	2.94 ^{ab} ±0.35	5.19 ^{bc} ±1.0	4.42 ^{bc} ±0.39	2.81 ^b ±0.15
Zalcium m)	88	3.92ª ±0.14	19.4**11.55*±3.2 ±3.5	8.71 ^{ab} ±1.58	3.9ª ±1.39	20.9 ^b ±4.37	5.75° ±0.88	10.5ab ±1.93	10.8ª +2.19	7.6ª ±0.87
Exchangeable Calcium (meq/100gm)	SI	7.18 ^{abc} ±1.52	19.4*c11 ±3.5	23.8abc ±4.22	5.3b ±1.24	27.4° ±3.72	9.6 ^{ab} +1.61	24.9abc ±3.96	18.9°c +2.39	13.2 ^{ab} ±0.99
	SS	0.58ns ±0.25	0.69°s ±0.17	0.98 ^{ns} ±0.28	0.42m ±0.14	0.1 ^{ns} ±0.25	0.53 ^m ±0.09	0.96m ±0.43	0.79ns ±0.12	0.47ms ±0.06
Exchangeable Potassium (meq/100gm)	TS	0.47 ^{ab} ±0.02	0.94 ^{ab} ±0.17	0.54 ^{ab} ±0.36	0.39° ±0.05	1.35 ^a ±0.3	0.82 ^{ab} ±0.13	1.35 ^{ab} ±0.5	1.21 ^a ±0.14	0.77ab ±0.08
eable m Ogm)	SS	0.36 ^{rs} ±0.05	0.37 ^{rs} ±0.06	0.41 ^{ns} ±0.11	1.85ns ±1.57	0.44 ^{ns} ±0.05	0.26 ^{ns} ±0.26	0.31 ^{ns} ±0.09	0.41 ^{ns} ±0.55	0.25 ^{rs} ±0.26
Exchangeable Sodium (meq/100gm)	SL (0.23ª ±0.00	0.28ª ±0.03	0.34ª ±0.08	0.18ª ±0.02	0.37a ±0.05	0.25a ±0.03	0.96 ^b ±0.52	0.36ª ±0.03	0.27a ±0.02
ole rrus	88	11.9°s ±2.5	89.6 ^{rs} ±48.3	11.9°s ±2.1	25.9°s ±6.3	45.2ns +22.1	14.4ns ±4.9	16.8°° ±13.3	5.6 ^{rs} ±1.1	15.8°s ±11.2
Available Phosphorus (ppm)	TS	26.3 ^{ab} ±9.1	59.2 ^b ±15.8	13.3 ^{ab} ±1.7	24.2 ^{ab} ±7.2	54.6 ^{ab} ±25.6	15.4 ^{ab} ±3.8	26.6 ^{ab} ±17.3	6.0ª ±1.6	4.2ª ±1.5
trogen	88	0.69ª ±0.08	0.23 ^b ±0.04	0.34 ^{bc} ±0.04	0.39° ±0.03	0.26bc ±0.02	0.31 ^{bc} ±0.04	0.21 ^{bc} ±0.02	0.29k ±0.01	0.2 ^b ±0.01
Comm. Total Nitrogen Type (%)	TS	0.9a ±0.08	0.51 ^{bc} ±0.03	0.58 ^{bc} ±0.05	0.62° ±0.07	0.48bc ±0.04	0.44bc ±0.04	0.44bc ±0.02	0.44bc ±0.03	0.38 ^b ±0.02
Comm. Type		1	7	භ	4	ľ.	9	L	∞	6

The major discrimination among the community types is due to altitude. Bonnefille *et al.* (1993) reported from palynological studies of forests and woodlands the presence of altitudinal zonation delimiting vegetation types in southwestern Ethiopia. Altitude is an important environmental factor that affects atmospheric pressure, moisture, and temperature which have a strong influence on the growth and development of plants and the distribution of vegetation (Hedberg, 1964).

The topsoils were categorized into 5 textural classes and the subsoils into 3: loamy sand (communities 1 and 4), clay loam (community 2), sandy loam (community 3), sandy clay loam (community 5), and loam (communities 6, 7, 8, and 9). The subsoil texture class includes sandy loam (communities 1 and 4), loam (community 3), and clay loam (communities 2, 5, 6, 7, 8, and 9). Soil texture is an important soil parameter that affects site quality. It influences the nutrient supplying ability of soil solids, soil moisture and air relations, and root development (Spurr and Barnes, 1980).

Comparison of the community types based on the sand content of the soil shows that type 1 is significantly different from types 2, 5, 6, 8, and 9, and type 4 differs from types 2, 5, 6, 7, 8, and 9 in both the topsoil and subsoil, and type 6 differs from types 3, 5, and 9 for topsoil. Community type 4 differs from types 8 and 9 in its silt content in the topsoil and from types 7 and 9 in the subsoil. Community types 8 and 9 differ in their silt content in the subsoil. Community type 2 differs from types 1 and 9 and community type 4 from type 2 and 6 in the clay content of topsoil and community types 4 and 8 differ in the clay content of subsoil.

The soils of southwestern Ethiopia are acidic with pH value ranges between 4 and 6. Acidic soils of medium intensity (*i.e.*, pH 4 and 5) were recorded for community types 1, 4, and 6 while slightly acidic soils (*i.e.*, pH 5 and 6) were recorded for the rest. The acidity of the soil could be caused by a more intense breakdown of organic matter and leaching of the soil attributable to the very high annual rainfall. Soil pH affects the growth of plants and the distribution of vegetation types by its effect on the availability of mineral nutrients and decomposition of organic matter (Buckman and Brady, 1969).

Community type 1 is different from the remaining types in soluble salts in both the topsoil and subsoil, and type 2 differs from type 9 in topsoil. The electrical conductivity of the soil solution is mainly determined by soluble salts of carbonates, bicarbonates, sulphates, chlorides and nitrates (Chopra and Kanwar, 1982). The concentration of salts in the soil solution influences an exchange of nutrients between the soil solution and plants.

The organic matter content of the soils is generally high (8.6–20.1%). There is significant variation in the organic matter content of the soils of the plant community types. Community type 1 is different from the other types and

type 4 differs from types 2, 5, 6, 8, and 9. Community type 3 differs from types 2, 8 and 9. Community types 1 and 4 are from a region which receives higher amounts of annual rainfall with lower temperature. The higher amount of organic matter is due to the inhibition of mineralization of organic matter by low temperature and acidic pH. The high organic matter content is important due to its impact on the fertility status and productivity of the soils. The mineralization of organic matter contributes to the supply of available mineral nutrients for plant use (Murphy, 1969). Higher amount of organic matter content has also been reported from the humid forest of Jibat (Tamrat Bekele, 1994). Westphal (1975) indicated that dark red-brown soils that are slightly to strongly acid have higher organic matter content.

The total nitrogen content of the soil follows closely the amount of organic matter. It is higher in those community types with higher organic matter content and lower in those having lower organic matter content. Much of the soil nitrogen is in the organic form and its distribution in the soil is approximately as that of the organic matter (Thompson and Troeh, 1979).

Community types 2, 4, and 5 have higher amount of available phosphorus, types 3, 6 and 9 have medium amount and type 8 has lower amount of available phosphorus both in topsoil and subsoil while community types 1 and 7 have higher amount in topsoil and medium amount in subsoil. There is no much difference between the community types in available phosphorus except the differences between community type 2 and types 8 and 9.

The soils of the community types are rich in exchangeable bases particularly in exchangeable calcium, which accounts for 70.5–79.9% and 54.5–76% of the exchangeable bases in the topsoil and subsoil, respectively. There is no variation among the community types in exchangeable sodium except for type 7, which differs from the other types for topsoil only. Community type 4 differs from types 5 and 8 in its exchangeable potassium and from types 2, 5, and 8 in exchangeable calcium of the topsoil. In the subsoil, type 5 differs from types 1, 2, 4, 6, 8, and 9 in exchangeable calcium. The variation could be accounted to the differences in soil pH and the amount of rainfall. The amount of exchangeable calcium and potassium present in the soil decline as a soil becomes more acidic and increases as the acidity declines (Thompson and Troeh, 1979). Community type 5 significantly differs in exchangeable magnesium from community types 1, 4, 6, and 9 in the topsoil and community types 1, 2, 4, 6, 8, and 9 in the subsoil and community type 4 differs from community types 2, 3, 5, 7, 8, and 9 in the topsoil.

The exchangeable bases have their highest concentration in community types 3, 5, and 7 and these communities occurred in areas that received the least amount of rainfall resulting in little loss of the cations by leaching.

The cation exchange capacity (CEC) of the soils is generally low. The CEC of the soils of community types 3, 4, and 5 is higher than the soils of types 2 and 9. The factors that influence CEC are texture, organic matter, and pH of the soil (Thompson and Troeh, 1979), but these contrasting groups of community types showed different degree of variability in terms of these factors, therefore, no single factor is responsible for the observed variation. The base saturation of the soils increases from 19.8% to 83.3% as pH rises from 4.1 to 5.8. Those community types whose soils' pH are relatively higher have most of their cation exchange portion occupied by basic cations (K⁺, Na⁺, Ca⁺⁺, and Mg⁺⁺) and those whose pH is lower have most of their cation exchange portion occupied by acidic cations (H⁺ and Al⁺⁺⁺) (Thompson and Troeh, 1979).

The results of Pearson's product-moment correlation of the environmental parameters (Table 3a and 3b) show that some of the environmental parameters are correlated. Altitude is positively correlated with sand and organic matter in topsoil and with electrical conductivity, organic matter and total nitrogen in subsoil, but negatively with clay in both topsoil and subsoil and available phosphorus in subsoil. Organic matter, total nitrogen and electrical conductivity are positively correlated in both topsoil and subsoil.

With increasing altitude, rainfall amount increases and this results in the loss of the silt fraction first, the clay fraction next and the sand fraction is the last to be washed away. The correlation of organic matter with electrical conductivity could be explained by its capacity to supply replaceable cations to the soil solution that increases the conductivity. The sand content of the soil is positively correlated with organic matter and total nitrogen but negatively with silt and clay content of the soil in both the topsoil and subsoil.

The correlation of organic matter and nitrogen with sand contradicts the fact that sandy soils usually carry less organic matter and nitrogen (Buckmann and Brady, 1969). pH is positively correlated with silt, exchangeable calcium, magnesium, and potassium (topsoil) and negatively with sand (subsoil), organic matter and total nitrogen. When the soil pH is lower the rate of organic matter decomposition decreases. This is due to the effect of soil pH on the activity of soil microrganisms that are involved in the decomposition of organic matter. The positive correlation between pH and calcium, magnesium, and potassium is to be expected as soils that are highly leached have lower pH because the basic cations have lower proportions than the acidic cations (Buckman and Brady, 1969). Among the exchangeable bases, potassium, calcium, and magnesium are positively correlated.

Table 3a. Pearson's product-moment correlation coefficient between environmental parameters (topsoil). *=p<0.05.

	Alt Slope Exposure Sand	re Sand Silt	lt Clay	Ηd	Clay pH Conductivity Organic matter	/ Organic matter	Z	Ъ	Na	<u>.</u>	У	ප	Mg) E
Alt	•				·									
Slope	-0.49 -													
Exposure	-0.06 -0.24 -													
Sand	0.73* -0.42 0.09	1,												
Silt	-0.47 0.05 0.32	-0.70* -												
Clay	-0.84 0.64 -0.28	-0.89* 0.39												
Hd	-0.66 0.39 0.46	-0.65 0.85*	5 0.50				,							
Conductivity	0.45 -0.44 -0.2	0.61 -0.54	4 -0.41 -0.62											
Organic matter	0.69* -0.18 -0.17	0.77 -0.8*	-0.57	-0.82* 0.78*	0.78*	•			•					
Z	0.65 -0.24 -0.19	0.79* -0.8*	-0.62	-0.75 0.82*	0.82*	.680						,		
Ь	-0.57 0.37 0.22	-0.04 -0.21	0.33	0.22	0.18	0.29	0.13	ı						
Na	-0.12 -0.46 0.85	-0.16 0.56	6 -0.12 0.48		-0.07	-0.37	-0.32	0.05	1					
⊻	-0.17 0.30 0.63	-0.38 0.59	0.18	0.74	-0.48	-0.37	-0.47	90.0	0.56	i				
ප	-0.35 0.31 0.70	-0.35 0.58	0.21	0.81	-0.42	-0.30	-0.45	0.34	09.0	,6·0				
Mg	-0.29 0.32 0.64	-0.45 0.61	0.26	080	-0.52	-0.40	-0.54	0.18	0.57	0.97		0.95		
CEC	0.66 0.07 0.25	0.63 -0.56	-0.59	-0.36 0.19	0.19	99.0	0.51	0.04	-0.07	7 0.24		-0.30	0.18	: (

	Alt Slope Exposure Sand	rre Sand Silt Clay pH Conductivity Organic matter	Organic matter	Z	<u>C</u> ι	Na	×	ඊ	Mg	CEC
Alt										
Slope	-0.49 -									
Exposure	-0.06 -0.24 -									
Sand	0.63 -0.14 -0.25						•			
Silt	-0.22 -0.29 0.46	-0.77								
Clay	-0.73* 0.47 -0.04	-0.81* 0.26 -								
Hd	-0.63 0.18 0.62	-0.81* 0.68* 0.61 -								
Conductivity	0.70* -0.34 -0.45	0.57 -0.36 -0.53 -0.65 -								
Org	0.78* -0.24 -0.31	0.82* -0.53 -0.76* -0.77* 0.89*			٤					
Z	0.74* -0.31 -0.36	0.78* -0.55 -0.67* -0.78* 0.94*								
o.	-0.77, 0.61 0.06	-0.08 -0.20 0.28 0.32 -0.40	-0.29	-0.30						
Na	0.20 -0.14 0.01	0.64 -0.63 -0.39 -0.49 -0.10	0.16 0.	0.18	0.02	·				
K	-0.06 0.18 0.72*	-0.36 0.31 0.26 0.66 -0.26	-0.21 -0	-0.28	90.0	-0.40				
ප	-0.54 0.41 0.40	-0.47 0.12 0.60 0.77 -0.42	-0.51 -0	-0.51	0.42	-0.31	0.72	•		
Mg	-0.48 0.43 0.43	-0.53 0.18 0.64 0.78 -0.48	-0.56 -0	-0.57	0:30	-0.33	0.78*	.86.0	i	
CEC	0.54 0.12 0.12	0.66 -0.63 -0.63 -0.35 0.20	0.46 0.	0.37	-0.18	0.46	0.30	0.14	0.17	

Phytogeographical comparison

The forests of southwestern Ethiopia are compared with other montane forests in Eastern and Southern Africa: two in Ethiopia, two in other East African countries and one in Southern Africa. The forests included in the comparison are Jibat forest and Harena forest in Ethiopia, Mt. Kenya forest in Kenya, Mt. Mulanje forest in Malawi and Mt. Elgon forest in Kenya/Uganda.

Jibat forest is located in western Shewa in western Ethiopia and lies at altitudes between 2000 and 3000 m (Tamrat Bekele, 1994). Harena forest is situated in the Bale Mountains National Park in southeastern Ethiopia with altitudinal distribution between 1510 and 3300 m (Lissanework Nigatu, 1987, Zerihun Woldu *et al.*, 1989). Mt. Mulanje, the highest massif in Malawi, is situated in the south-east part of Malawi. The 'mid-altitude' and 'Afromontane forests' *sensu* Dowsett-Lemaire (1988) of Mt. Mulanje are distributed between 900 and 1950 m a.s.l. Mt. Elgon is a large volcano of Mioecene age straddling the Kenya/Uganda border. A belt of forest girdles the mountain between 2000 and 3000 m a.s.l. (Hamilton and Perott, 1981). Mt. Kenya is an isolated mountain of volcanic origin in Kenya. Forests cover the bulk of the mountain area, extending to approximately 3400 m altitude in the south and 3000 m in the north (Bussmann and Beck, 1995).

The comparison is based on the similarities in species distribution. A similarity analysis was carried out based on the presence of tree and shrub species in order to evaluate the relationship between these forests. The similarity index used is Sörensen's similarity coefficient 2c/(a+b), where c is the number of species shared by the forests compared, a is the number of species in one forest, and b is the number of species in the other forest. The results of the analysis are presented in Table 4.

Table 4. Floristic similarities between the forests in southwestern Ethiopia and five forests in Ethiopia, Kenya, Kenya/ Uganda and Malawi. N = number of species included in comparison, C = number of species in common, S = Sörensen's coefficient of similarity.

Forest	N	С	S
Jibat	54	39	0.54
Harena	87	51	0.58
Mt. Mulanje	137	30	0.26
Mt. Elgon	61	22	0.29
Mt. Kenya	160	39	0.3

The forests in southwestern Ethiopia are floristically related more to the Harena and Jibat forests than to the forests on Mt. Kenya, Mt. Mulanje and Mt. Elgon. The Harena and Jibat forests are situated in the same climatic region as

those of the forests in southwestern Ethiopia. The Harena forest in the moister side of the Southeastern Highland and the Jibat forest in the moister side of the Northwestern Highland. Therefore, the highest floristic similarity between these forests is due to similar distribution in the moister parts of the country. The forests of Mt. Kenya, Mt. Mulanje and Mt. Elgon are exposed to a moister climate but they are geographically far from the forests of southwestern Ethiopia.

Hedberg (1951) recognized three-vegetation belts in East African mountains: Montane belt; Ericaceous belt; and Afroalpine belt. The montane belt is divided into three zones: montane forest zone; bamboo zone; and *Hagenia-Hypericum* zone. The Ericaceous and Afroalpine belts are not represented in southwestern Ethiopia, only the montane belt is represented. The montane forest and bamboo zones have been described in this study; the former from Jiren, Yayu, Gebre Dima, Sele-Anderacha, Bonga, and Belete-Gera forests and the latter from Sele-Anderacha forest. The *Hagenia-Hypericum* zone has not been encountered, this is because the altitudinal range covered in this study is below the lower limit of *Hagenia-Hypericum* zone.

Among the seven forest types described by Friis (1992), four are represented in southwestern Ethiopia: Afromontane rain forest; Transitional rain forest; Dry peripheral semi-deciduous Guineo-Congolian forest; and Riverine forest. Except Sheko forest, the remaining forests investigated in this study are Afromontane type with characteristic Afromontane flora.

Among the Afromontane taxa some are endemic, i.e., confined to the Afromontane region only, e.g. Pouteria adolfi-friederici, Dracaena afromontana, Millettia ferruginea (Hochst.) Bak, Olea welwitschii, and Arundinaria alpina. Some Afromontane endemic species are forest pioneers, e.g., Bersama abyssinica Fresen. Some others are linking species, i.e. occurring in the Afromontane region and in other phytochoria, e.g. Croton macrostchyus, Ficus thonningi Blume, and Phoenix reclinata Jacq.. Other species are transgressors, i.e., occurring in two or more major phytochoria and are characteristic members of two or more vegetation types, e.g., Cassipourea malosana (Bak.) Alston and Celtis africana.

The Transitional rain forest type is represented in Sheko forest and in the river valleys of Yayu forest. The species in these forests are either characteristic Afromontane, (e.g. *Manilkara butugi, Olea hochstetterri* Bak., and *Macaranga capensis* (Baill.) Sim) or Guineo-Congolioan, (e.g. *Pouteria altissima* and *Argomuellera macrophylla*) or Guineo-Congolian linking species, (e.g. *Alchornea laxiflora, Celtis zenkeri, Dracaena fragrans, Ehretia cymosa* Thonn., *Morus mesozygia, Rothmannia urcelliformis* (Hiern) Robyns, *Sapium ellipticum* (Hochst. ex Krauss.) Pax, and *Trilepisium madagascariense* DC.

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